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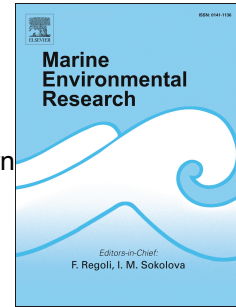
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1 **DNA metabarcoding to assess prey overlap between tuna and seabirds in the Eastern Tropical**
2 **Atlantic: implications for an ecosystem-based management**

3

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23

24 Abstract

25 Overfishing has been drastically changing food webs in marine ecosystems, and it is pivotal to
26 quantify these changes at the ecosystem level. This is especially important for ecosystems with a high
27 diversity of top predators such as the Eastern Atlantic marine region. In this work we used high-
28 throughput sequencing methods to describe the diet of the two most abundant tuna species, the
29 Skipjack tuna (*Katsuwonus pelamis*) and the Yellowfin tuna (*Thunnus albacares*), highly targeted by
30 fisheries off west Africa. We also explored prey diversity overlap between these tuna species and the
31 seabird species breeding in Cabo Verde that are most likely to share prey preferences and suffer from
32 bycatch, the Brown booby (*Sula leucogaster*) and Cape Verde shearwater (*Calonectris edwardsii*).
33 Overall, the diet of both tuna species was more diverse than that of seabirds. Skipjack tuna diet was
34 dominated by prey from lower trophic levels, such as krill, anchovies, and siphonophores, while the
35 Yellowfin tuna diet was mainly based on epipelagic fish such as flying and halfbeak fishes. Some of
36 the most abundant prey families detected in the Yellowfin tuna diet were shared with both seabird
37 species, resulting in a high prey diversity overlap between this tuna species and seabirds. These results
38 have implications for the management of tuna fisheries in the Eastern Tropical Atlantic, because a
39 large decrease of both tuna species might have cascading effects on both primary and secondary
40 consumer levels, and the decrease of these underwater predators may have implications on the
41 viability of tropical seabird populations.

42 **Keywords:** Diet, DNA metabarcoding, Tropical ecosystem, *Katsuwonus pelamis*, *Thunnus*
43 *albacares*, *Sula leucogaster*, *Calonectris edwardsii*, Cabo Verde

44

45

46 **1. Introduction**

47 Overfishing around the globe has been drastically reducing populations of target and non-target
48 marine species, with major effects on ecosystem structure and food web functions (Jennings et al.,
49 2001; Pauly et al., 1998). These effects need to be quantified under an ecosystem-based approach,
50 especially incorporating top predators, such as tunas (also overfished) and seabirds (Fowler, 1999;
51 Karpouzi et al., 2007; Piatt et al., 2007). As predators on the top of marine food chains, seabirds and
52 tunas have a major role in food web structure at the ecosystem level, because they contribute to the
53 stability and persistence of marine ecosystems through top-down control (Estes et al., 2011).
54 Information about their ecological niche and trophic dynamics is therefore necessary, for instance, to
55 understand and predict potential cascading effects caused by the decrease of large predatory fishes
56 due to overfishing (Cairns, 1992; Myers et al., 2007).

57 This is especially important for highly diverse ecosystems of tropical waters. On one hand, these
58 ecosystems harbour a great diversity of both aerial (such as seabirds) and underwater (such as tunas,
59 dolphins and sharks) top predators. On the other hand, these waters are usually less productive than
60 their temperate counterparts, with scarcer and more patchily distributed prey for top predators
61 (Longhurst and Pauly, 1987). To overcome this, predators tend to adopt foraging strategies which
62 allow the detection of these prey patches, whilst minimizing energy spent foraging. For instance,
63 seabirds can adopt foraging strategies that take into account sub-surface predators, especially tunas
64 (Spear et al., 2007). More specifically, tunas often chase prey to the ocean surface, which facilitates
65 prey availability for seabirds (Ashmole and Ashmole, 1967). This kind of facilitated foraging has not
66 only been directly observed in the environment (e.g. Au & Pitman 1986, Hebshi et al. 2008, Correia
67 et al. 2019), but has also been inferred from the overlap between seabirds' and aquatic predators'
68 foraging areas (e.g., Catry et al. 2009), of their trophic niches (e.g., Kojadinovic et al. 2008) and prey
69 species (e.g., Ménard et al. 2012). For example, in the Eastern Tropical Pacific facilitated foraging is
70 the single most important strategy for seabirds' successful foraging, accounting for a total of 76% of
71 prey mass made available to seabirds (Spear et al. 2007). It seems that these interactions might also

72 be beneficial for tuna because seabirds might prevent prey such as flying fish from leaping out of the
73 water, containing the prey at the surface (Burger 1988). Furthermore, it is expected that climate
74 change may lead to the expansion of ocean hypoxic zones, which may limit pelagic tuna to the thinner
75 oxygenated surface layer (Stramma et al., 2012; Vedor et al., 2021), and thus such interactions with
76 seabirds might be more beneficial than expected (although competition for food between tuna and
77 seabirds might also arise from such interactions). Several studies have shown, however, that while
78 tunas and seabirds forage in the same schools, the two groups tend to forage at different trophic levels,
79 on different prey or different prey sizes (Ashmole and Ashmole, 1967; Bertrand et al., 2002b; Cherel
80 et al., 2008; Young et al., 2010). Therefore, the decrease of tunas on tropical waters due to overfishing
81 may not benefit seabirds due to competition release, instead, it may have the opposite negative effect
82 if seabirds' prey capture efficiency is intrinsically linked with tunas' capture events.

83 The East Atlantic marine region centred around Cabo Verde archipelago (Fig. 1a) is an iconic
84 example where overfishing may have a detrimental effect due to the large diversity of breeding
85 seabirds and the presence of several tuna species. However, little data is available on the sub-surface
86 predators' direct and indirect trophic interactions with seabirds. Indeed, the most captured sub-surface
87 predators in these waters are tuna species, especially Skipjack tuna (*Katsuwonus pelamis*), Yellowfin
88 tuna (*Thunnus albacares*) and Bigeye tuna (*Thunnus obesus*) (Monteiro, 2012; Thorpe et al., 2022).
89 On the other hand, Cabo Verde hosts eight breeding seabird species (including four endemic species)
90 with large colonies all over the archipelago (Semedo et al., 2021).

91 In Cabo Verde, the contribution of fisheries to the Gross Domestic Product (GDP) is relatively
92 small (Almeida et al., 2004; FAO, 2022), but they play a strategic social and economic role,
93 employing approximately 5.2% of the economically active population (González and Tariche, 2009;
94 Silva, 2009). Tuna species are the main target, comprising 30.5% of the total artisanal catches and
95 65% of the industrial fisheries (INE-CV, 2017). Favourable sea conditions allow fishing activity all
96 year round in Cabo Verde, however, the pressure due to unreasonable fishing quotas and the high
97 level of Illegal Unregulated and Unreported (IUU) catches are concerning (Ramos and Grémillet,

98 2013). Understanding resource partitioning between the most caught tuna species and seabird species
99 of Cabo Verde (Fig. 1b) will allow a better understanding of the effects that overfishing might have,
100 because their trophic ecology might reflect changes in the condition and availability of prey stocks
101 (e.g., Barrett & Krasnov 1996) or serve as indicators of changes in the ecosystem (e.g. Weimerskirch
102 et al. 2003).

103 Several methods can be used to quantify the diet of both tunas and seabirds, and analysis of
104 stomach contents has been used as a standard and reliable method for this purpose. However, the
105 morphological identification of stomach contents is time-consuming, and most importantly, can be
106 limited by the differential digestion of prey, with species composed of soft tissues often being missed,
107 and thus greatly underestimated. Also, the retrieval of stomach contents from seabirds through
108 stomach flushing, which involves pumping salt water through the animal's oesophagus (Barnett et al.,
109 2010; Wilson, 1984), can be time-consuming and is considered an invasive method (Harris and
110 Wanless, 1993). As an alternative, genetic-based tools on the stomach, regurgitates or faecal samples
111 can be applied. Metabarcoding combines genetic sequence-based identification with high-throughput
112 sequencing technology (HTS), which allows the identification of a broad range of taxonomic levels
113 at the same time, even from very degraded samples (Pompanon et al. 2012). The quantification of the
114 relative proportion of prey consumed using DNA Metabarcoding approach, on the other hand, is still
115 not achievable, and is considered the main limitation of this method for dietary studies (reviewed in
116 Ceia et al., 2022). Nonetheless, this method has proven to surpass morphological identification in
117 diverse species, including seabirds (e.g., Oehm et al. 2017, Xavier et al. 2018) and fishes (e.g.,
118 Kodama et al. 2017, 2020, Matley et al. 2018, Günther et al. 2021, Trujillo-González et al. 2022). It
119 has the additional advantage of minimizing disturbance to seabirds since it can be applied to
120 regurgitates or faecal samples.

121 In this work we described prey diversity of the most captured tuna species in Cabo Verde
122 using metabarcoding and explored prey diversity overlap with seabird species breeding in the
123 archipelago, which are most likely to interact with fisheries and potentially suffer bycatch (Montrond,

124 2020): the Brown boobies (*Sula leucogaster*) and Cape Verde shearwaters (*Calonectris edwardsii*).
125 This also allowed us to critically evaluate the likelihood of tuna species to act as facilitators of prey
126 availability to seabirds in Cabo Verde, and thereby contribute to a better understanding of this trophic
127 network. Ultimately, this work will also allow us to understand how prey depletion by fisheries could
128 trigger indirect trophic cascading effects and thus contribute to a better fisheries management in this
129 region.

130

131 **2. Materials and Methods**

132 2.1. Tunas sample collection and processing

133 With the collaboration of local fish processing plants and authorities, we collected stomachs of
134 the most representative underwater predatory species caught in fishing activities, the Skipjack
135 (*Katsuwonus pelamis*), Yellowfin (*Thunnus albacares*) and Bigeye (*Thunnus obesus*) tuna (Monteiro,
136 2012; Thorpe et al., 2022). All the sampled stomachs belonged to tunas captured between July and
137 early September of 2019, mostly by purse seine. Detailed data from all individuals, including the
138 fishing method used, fishing vessel, and date of capture is described in the Supplementary Data A.
139 Stomachs were immediately frozen at capture and preserved at a fish processing plant at the landing
140 harbour of Mindelo (Cabo Verde). In October 2019, also in Mindelo, all stomachs were processed in
141 laboratory conditions, in the Universidade Técnica do Atlântico (UTA). In more detail, each stomach
142 was thawed in individualized sterilized trays at room temperature for two to three hours or in the
143 fridge (4°C) overnight. From each individual tissue samples for barcoding genetic identification of
144 each tuna sample were also taken from other organs available (e.g., liver or heart) and stored in 2 mL
145 tubes with 96% ethanol. Since incorrect assignment of the tuna species through morphology might
146 occur at capture, this was done to genetically confirm the identification of the tuna species (further
147 described in section 2.4). These misassignments can be related either to the presence of many related
148 species on one single capture, or the ambiguous regional common names or fisherman's slang that
149 might encompass several species (Cawthorn and Mariani, 2017). Then each stomach was dissected,

150 its content sampled, collected into sterile Petri dishes, homogenized and two subsamples per stomach
151 were taken into 2 mL tubes with 96% (v/v) ethanol. All materials used between dissections were
152 cleaned in running water, then sterilized using bleach and 96% (v/v) ethanol to avoid cross-
153 contamination between samples. All samples were then refrigerated at 4°C as soon as possible,
154 transported to CIBIO research centre in Portugal and then stored at -20°C in the laboratory until DNA
155 extraction.

156

157 2.2. Seabird sample collection and processing

158 In 2018 and 2019, Brown boobies (*Sula leucogaster*) and Cape Verde shearwaters (*Calonectris*
159 *edwardsii*) were sampled in the Raso Islet (16°37'5" N, 24°35'15" W) and Boavista Island (15°59'15"
160 N, 22°47'08" W, Fig. 1a) during the chick-rearing period, from July to September, to overlap with
161 the tunas sampling period. These seabird species were selected based on the highest probability to
162 suffer from direct mortality through bycatch in Cabo Verde waters (Montrond, 2020). Each individual
163 was caught, processed, and released within 15 minutes. A faecal sample was collected from each bird
164 directly into a 2 mL tube with 96% (v/v) ethanol, after an abdominal massage to facilitate defecation.
165 These samples were refrigerated at 4°C as soon as possible, transported to CIBIO research centre in
166 Portugal and then stored at -20°C in the laboratory until DNA extraction.

167

168 2.3. Sequencing tuna tissue samples

169 Tuna muscle tissue samples undergone DNA extraction using the EZ-10 Spin Column DNA Gel
170 Extraction Kit protocol (Bio Basic Inc., USA). A 645 bp fragment of the mtDNA cytochrome c
171 oxidase subunit I (COI) was amplified with PCR using the primers LCOI 121 and HCOI 1199 (Paine
172 et al., 2007). The PCR protocol consisted of 5 µL of QIAGEN Multiplex PCR Master Mix (Qiagen,
173 USA), 0.4 µL of each primer (10 nM), 2 µL of DNA template and nuclease-free water for a total
174 volume of 11 µL. The PCR conditions were 15 min at 95°C, 40 cycles of 30 s at 95°C, 30 s at 57°C
175 and 60 s at 72°C, followed by 10 min at 72°C. All PCR products were sequenced on a 3730xl Genetic

176 Analyzer (Applied Biosystems, USA), and results were edited and assembled in Geneious Prime
177 2022.0.2 (Biomatters, New Zealand).

178 Species assignment was validated using both Phylogenetic tree and BLAST search for more
179 robust results. First, for each mtDNA region, sequences were aligned and an approximately-
180 maximum-likelihood algorithm was used to build a phylogenetic tree in FastTree 2 (Price et al., 2010)
181 to taxonomically assign our sequences. For this approach, we used reference sequences for all species
182 of tuna and related species known to occur in the study area. Second, using the BLAST algorithm
183 (Ye et al., 2006), each sequence was compared (1,000 hits) against the NCBI Genbank Nucleotide
184 online data repository to further confirm the assignment of each sequence to a single species.

185 We found that a miss-assignment occurred in 33 % (19 out of 57) of the individuals. This did
186 not impact the sample size for Skipjack (final n = 28) and Yellowfin (final n = 23) but drastically
187 decreased our sample size for Bigeye (final n = 6), which prevented a robust statistical analysis of
188 Bigeye metabarcoding data. Thus, the present study focused on the statistical analysis of the other
189 two species, and data on metabarcoding on Bigeye tuna is made available in Supplementary Data B
190 as valuable data for future research.

191

192 2.4. Diet determination

193 To avoid contaminations, all samples were extracted in a non-invasive laboratory using the Stool
194 DNA Isolation Kit (Norgen Biotek, Canada), following the manufacturer's protocol. For each batch
195 of samples to be extracted, a negative control prepared with distilled water was included. We chose
196 to amplify two different mitochondrial DNA (mtDNA) fragments not only to duplicate our probability
197 of correctly assigning the taxonomical identity of Osteichthyes – bony fishes prey items, but also to
198 cover a wider taxonomical range of potential prey (especially macro-invertebrates). The first set,
199 'MiFish-U' amplifies the 12S fragment, preferentially from Osteichthyes (Table 1). The second set,
200 mlCOIintF-XT and jgHCO2198, referred to as 'Leray-XT' hereafter, amplifies the COI fragment
201 from metazoan sources, including therefore not only fish but also cephalopods and crustaceans (Table

202 1). Preliminary analysis of a small subset of seabird and tuna samples showed that both primer sets
203 amplified the predator's DNA. Tuna's stomach content samples can contain tuna blood, and thus high
204 concentrations of predator's DNA. Therefore, blocking primers had to be designed for decreasing the
205 amplification of Tuna DNA in these samples (Table 1). On the other hand, seabird's faecal samples
206 also may have epithelial cells from the predator digestive system, but these preliminary tests proved
207 that blocking primers were not necessary for these faecal samples since they contained less predator
208 DNA, and enough reads from prey items (more than 50%) were always obtained in these samples.

209 Library preparation followed the MiSeq protocol for 16S Metagenomics (Illumina, USA). For
210 tuna samples, four libraries were prepared, three with 'MiFish-U' primer set and one with 'Leray-
211 XT'. For the 'MiFish-U' the three libraries prepared were 1) without blocking primer, 2) with
212 blocking primer at 20x, and 3) with blocking primer at 100x. The 'Leray-XT' library was prepared
213 with 100x blocking primer. For seabirds' samples, two libraries were prepared, one for each primer
214 set without blocking primer. For each sample, PCR reactions were carried out in volumes of 10 to 12
215 μl , comprising 5 μl of Multiplex PCR Master Mix (QIAGEN, Valencia, CA, USA), 0.2 μl of each
216 forward and reverse primers (10 nM), the proper volume of the blocking primer depending on the
217 library (100 nM), 1 μl of DNA extract and filled the left-over volume with ultra-pure water. All PCRs
218 included two negative controls to test for possible contamination, one from the DNA extraction
219 process, as well as a PCR negative control prepared with distilled water. PCR cycling conditions are
220 described in detail in Appendix A.

221 Amplification success was checked by visually inspecting 2 μl of each PCR product on a 2%
222 gel-stained agarose (GelRed, Biotium, USA). Then, for individual identification of each amplified
223 product before pooling, the PCR products were amplified again in a PCR with P5 and P7 indexes,
224 each containing a unique 7 bp long barcode that differed at least 3 bp from any other index. PCR
225 reactions and cycling conditions were similar to the previous PCR, however, only 10 cycles of
226 denaturing, annealing and extension were done, with an annealing temperature of 55°C. Indexing
227 success was also verified on gel-stained agarose, then indexed PCR products were purified using

228 Agencourt AMPure XP beads (Beckman Coulter, USA), and subsequently quantified using an Epoch
229 Microplate Spectrophotometer (BioTek, USA). All PCR products from each marker were pooled and
230 normalized at equimolar concentrations (15 nM). Pools were then purified again and quantified into
231 libraries using qPCR with a KAPA Library Quant Kit qPCR Mix (KAPA Biosystems, USA) on the
232 iCycler Real-Time PCR Detection System (Bio-Rad, USA), and further diluted to 4 nM. The final
233 library was run in a MiSeq sequencer (Illumina) using a v3 MiSeq reagent kit (Illumina) for an
234 expected average of 24,000 paired-end reads per sample.

235

236 2.5. Bioinformatic analysis

237 First, paired-end reads were aligned using PEAR (Zhang et al., 2014), discarding both
238 unassembled reads and alignments with overlapping quality scores < 26 . Further processing of
239 sequencing reads was done using OBITools (Boyer et al., 2016), where reads were assigned to
240 samples and primer sequences were removed using ‘ngsfilter’, allowing a total of four mismatches.
241 Using ‘obiuniq’ reads were collapsed into amplicon sequence variants (ASVs) and error-corrected
242 using ‘cluster_unoise’ from VSEARCH (Rognes et al., 2016). Only for the ASVs obtained from the
243 COI fragment, additionally we used VSEARCH to perform a 99% identity clustering step into
244 Molecular Operational Taxonomic Units (MOTUs). The retained sequences were further filtered
245 using LULU (Frøslev et al., 2017) by discarding sequences with over 84% similarity and 95% co-
246 occurrence for ‘Leray-XT’, and 95% similarity and 95% co-occurrence for ‘MiFish-U’. Finally, a
247 further filtering process excluded PCR products that exhibited less than 100 reads in total from the
248 analysis, as well as all ASVs/MOTUs (hereafter all named MOTU for the sake of simplicity) that had
249 a read count $< 1\%$ of the total number of reads of that PCR (Mata et al., 2019).

250 Prey items were identified by comparing the final MOTUs against online databases (BOLD and
251 NCBI Nucleotide Database) using BLAST algorithms. Haplotypes were assigned to the lowest
252 possible taxonomic level (e.g., family, order, species) for which 1000 hits in BLAST, with the highest
253 identity and identical match, clustered monophyletically. A curation step was performed to assess if

254 each MOTU was described as occurring in the study area, consulting online databases such as GBIF
255 (GBIF.org, 2022) or FishBase (Froese and Pauly, 2022). Identifications that failed this curation step
256 and also were not found to be present on the Eastern Atlantic Ocean using the same online databases
257 were conservatively assigned to the following taxonomic level. Finally, every MOTU that belonged
258 to other taxa than the possible prey species (e.g., human, other mammals, parasites or bacteria) were
259 discarded and a final matrix of MOTUs per sample was built.

260

261 2.6. Data Analysis

262 After bioinformatic filtering, an average of 6876 diet reads per sample was obtained and a total
263 of 48 tuna samples (20 of yellowfin tuna, 28 of skipjack tuna) and 43 seabird samples (21 for Brown
264 boobies and 22 for Cape Verde Shearwaters) were used for analysis. From the three ‘MiFish-U’
265 libraries, and to avoid overrepresentation over the ‘Leray-XT’ library, for each sample it was chosen
266 the library that presented the highest proportion of diet DNA, which in most cases was the library
267 built with blocking primer at a concentration of 100 x (Appendix B). Data was shown at the MOTU
268 and family level and inferential analysis was done only at the family level. Our rationale is that many
269 prey MOTUs will share the same trophic biologically relevant phenotypic characteristics (behaviour,
270 size/energy ratios) and the analysis of MOTUs may have a lower statistical power due to the high
271 number of MOTUs detected between tunas and seabirds (79 MOTUs). All MOTU-level
272 visualizations can be consulted in Supplementary Data C.

273 For each predator species, and independently of the primer set for which it was amplified, the
274 frequency of occurrence (%FO) of each MOTU and MOTU family was estimated and ranked %FO
275 curves were plotted. The occurrence of MOTU and MOTU families between tunas and seabirds was
276 visualized using the ‘upset’ command from the *UpSetR* package (Conway et al., 2017). To visualize
277 MOTU and MOTU family richness variation taking into account sample size, for each of the four
278 predators, an accumulation curve was calculated at both levels using ‘specaccum’ command from the
279 *vegan* package (Oksanen et al., 2020). Observed and extrapolated MOTU and MOTU family richness

280 was calculated using ‘iNEXT’ command from the *iNEXT* package (Hsieh et al., 2020). The
281 extrapolated richness and its standard error were calculated for an endpoint of twice the sample size
282 of the predator species with the lower sample size. We compared MOTU family observed richness
283 estimates between both tuna and seabird species using a chi-squared goodness-of-fit test followed by
284 pairwise comparisons, calculated using the ‘chisq.multcomp’ command from the *RVAideMemoire*
285 package (Herv, 2022).

286 The Pianka niche overlap index (O_{jk} , Pianka 1973) was calculated at the family level to
287 understand the niche overlap between the prey diversity of the four species using ‘niche.overlap’
288 command from the *spaa* package (Zhang and Ma, 2014). Null models were used to test whether the
289 extent of niche overlap is greater than expected by chance, where 1,000 simulated MOTU matrices
290 were built using ‘niche_null_model’ with the ‘ra3’ algorithm, from *EcoSimR* package (Gotelli and
291 Ellison, 2013). The observed niche overlap was then compared to the simulated niche overlap values,
292 which were considered significant when the observed value was greater than 95% of the simulated
293 values ($p < 0.05$).

294 To explore differences in diet composition, a multivariate analysis of the MOTU family
295 occurrence matrix was performed with tunas and seabirds’ data, to test for interspecific (Species
296 factor) and temporal (Month factor: June to September) differences. Generalized linear models for
297 multivariate presence/absence data were fitted using the *mvabund* package (Wang et al., 2012) with
298 the ‘manyglm’ command. Since seabird samples were collected in two different years, interannual
299 differences in the diet composition of seabirds were tested prior to this analysis. This factor was found
300 to be not significant, and thus we continued the analysis with samples from both years to obtain a
301 more balanced sample size for each predator species. To test for each factor's significance, as well as
302 their interactions, a model-selection approach was performed using the Akaike Information Criterion
303 (AIC) and deviance using the ‘anova.manyglm’ command. The p-values for the effect of each
304 variable in the alternative model were obtained by Bootstrap resampling (1000 x) of a log-likelihood
305 ratio under the null model. To understand which MOTU families expressed significant effects,

306 univariate statistics were calculated, and resampling-based univariate p-values were obtained.
307 Finally, a non-metric multidimensional scaling (nMDS) was used to visualize diet composition
308 dissimilarities, built with the function ‘metaMDS’, also from the *vegan* package. All analyses and
309 visualizations were made under the statistical environment R 4.1.3 (R Core Team, 2022).

310

311 **3. Results**

312 The ‘MiFish-U’ identified MOTUs families in most of the tuna samples (38 out of 48) and also
313 in the seabird samples (38 out of 43). In the remaining samples, MOTUs families were only amplified
314 with the ‘Leray-XT’. Both sets of primers identified MOTUs families in 21 and 8 samples from tunas
315 and seabirds, respectively. The use of the designed blocking primers notably reduced the proportion
316 of tuna DNA, enhancing the proportion of diet DNA per sample (Appendix B). However, the
317 designed blocking primers were more effective on Yellowfin and Bigeye tuna samples than in
318 Skipjack tuna, especially at the concentration of 100X.

319 Overall, both tuna species preyed on a higher number of families in comparison with seabirds
320 (Fig. 2, Appendix C). Both tunas presented the highest family richness, both observed (Yellowfin =
321 30; Skipjack = 20) and extrapolated (Yellowfin = 47.04 ± 12.42 ; Skipjack = 25.66 ± 6.94). Seabirds
322 showed the lowest observed family richness (Cape Verde shearwater = 14, Brown booby = 10),
323 however, Cape Verde shearwater presented similar extrapolated family richness to Skipjack ($21.00 \pm$
324 10.72), while Brown booby presented the lowest values (12.44 ± 4.66). These differences in the
325 overall observed family richness were significant ($\chi^2 = 12.27$, $df = 3$, $p = 0.007$), where Yellowfin
326 tuna family richness was significantly higher than Cape Verde shearwater ($p = 0.048$) and Brown
327 booby ($p = 0.009$). Even at the same sampling completeness (extrapolated family richness at $n = 40$),
328 the same tendency was found, with significant differences found in the overall extrapolated family
329 richness ($\chi^2 = 27.75$, $df = 3$, $p < 0.001$), where Yellowfin tuna family richness was significantly higher
330 than all the other predators (all $p < 0.009$).

331 A total of 12 families were shared between tuna species (Fig. 2), including 10 fish families, one
332 decapod family, and one krill family. Skipjack tuna diet was mainly dominated by three families (Fig.
333 3a), Euphausiidae (53.57% FO), Engraulidae (42.86% FO) and Diphyidae (28.57% FO), while the
334 other 17 prey families were present in much lower frequencies of occurrence (equal or lower than
335 17.86%). Yellowfin tuna presented a more generalist diet, with a higher number of prey families but
336 all with a low frequency of occurrence (Fig. 3a). The prey families with the highest frequencies of
337 occurrence were Hemiramphidae (30% FO) and Exocoetidae (25% FO).

338 From the total of prey families detected on the diet of both seabirds (17), four were shared
339 between the two predator groups: Exocoetidae, Hemiramphidae, Coryphaenidae and Serranidae (Fig.
340 2). Other five families were shared between at least one species of each predator group. Only five
341 prey families were exclusively detected in seabirds' diet: Congridae, Derichthyidae, Mullidae,
342 Muraenidae and Scombridae. The overlap index between seabird and tuna diets (Fig. 3b) was
343 significant between Yellowfin tuna and Brown booby ($p = 0.004$), and between both seabird species
344 ($p = 0.022$). The overlap index between Yellowfin tuna and Cape Verde shearwater, as well as
345 between Skipjack and both seabird species was not significant (all $p > 0.127$).

346 The multivariate data model analysis showed significant differences in diet composition between
347 all species (Species factor: $p = 0.001$) as well as significant temporal differences (Month factor: $p =$
348 0.003). The interspecific differences could also be visualized on the nMDS plot by the low overlap
349 and distance of centroids (Fig. 4), for which the multivariate data model showed the significant
350 contribution of six prey families (Table 2). Temporal differences were mainly explained by the
351 contribution of Gonostomatidae and Phosichthyidae families in August, when compared with the
352 other two months (Table 2, Supplementary data D). The interaction between both factors was also
353 significant ($p = 0.001$), mainly due to the differences between Skipjack tuna samples from August
354 with the other groups, with Carangidae and Exocoetidae contributing significantly to these differences
355 (Table 2, Supplementary data D).

356

357 **4. Discussion**

358 4.1. Tunas prey diversity

359 Both tuna species presented a diet with high species diversity, in line with the consensual view
360 of tunas as generalists and opportunistic predators (e.g., Ménard et al. 2006, Romero et al. 2021).
361 Yellowfin tuna exhibited a higher prey diversity than Skipjack tuna, corroborating previous studies
362 (Setyadji et al., 2012; Trujillo-González et al., 2022) which describe Yellowfin tuna as generalists,
363 also presenting a general low abundance of each prey type (Olson et al., 2014). This can be explained
364 by the fact that true tuna species have high metabolic rates and require higher energy intake for
365 continuous swimming activity (Magnuson, 1978; Olson and Boggs, 1986). They are also limited by
366 oxygen levels, due to their physiological adaptations, especially concerning heart rate (Pecoraro et
367 al., 2017; Stramma et al., 2012). With the increasing deoxygenation levels in tropical waters, prey
368 availability is further limited to a smaller range of depths. Since prey tend also to present a patchy
369 distribution at tropical latitudes, they must adopt a generalist, more opportunistic diet to ensure this
370 energy intake.

371 On the other hand, Skipjack tuna also presented high prey diversity, but some prey families
372 were more dominant, particularly krill (Euphausiidae), anchovies (Engraulidae) and siphonophores
373 (Diphyidae). Most studies on tropical waters describe the Skipjack tuna diet as mainly composed by
374 fish of the families Acanthuridae, Clupeidae, Scombridae, Carangidae and Serranidae (e.g. Dragovich
375 & Potthoff 1972, Setyadji et al. 2012, Romero et al. 2021). Other studies have described other prey
376 of lower trophic levels as the main components of Skipjack tuna's diet, namely krill (Alatorre-
377 Ramirez et al. 2017) and anchovies (Varela et al., 2019). These differences between Skipjack and
378 Yellowfin tuna diet composition can be mainly explained by differences in their body size. Not only
379 the smaller size and different physiology of Skipjack may limit the maximum prey size it can consume
380 (Graham et al., 2007), but also its anatomically smaller gill raker apparatus might retain a higher
381 abundance of small prey, such as euphausiids (Ankenbrandt, 1984; Magnuson and Heitz, 1971).
382 Because these two tuna species are present in these waters in higher numbers compared with other

383 tuna species (Monteiro, 2012), such consumption of prey of different sizes might reduce competition
384 and facilitate coexistence. From a trophic perspective, since DNA metabarcoding does not allow a
385 quantification of the relative proportion of prey consumed, and considering that our results are at the
386 family level, it is not possible to calculate a numerical average of the trophic levels of both tuna prey
387 diversity. However, both krill and anchovies, i.e., main Skipjack prey, are widely known prey of
388 lower trophic levels, since both families are comprised mainly of planktivorous species (Cabrol et al.,
389 2019; Chouvelon et al., 2014; Cleary et al., 2012; Quetin and Ross, 1991). In the other hand, some of
390 the most frequent prey of Yellowfin tuna are omnivorous fish which forage on different trophic levels,
391 such as fish from the Hemiramphidae family (Collette, 2016), or squid from the Ommastrephidae
392 family (Lipiński and Linkowski, 1988; Merten et al., 2017). Our results show that these two tuna
393 species forage on prey of different trophic levels, and such trophic level differentiation between
394 Yellowfin and Skipjack tuna was also observed in the tropical Pacific Ocean (Alatorre-Ramirez et
395 al., 2017).

396 Differences in diet in relation to earlier studies may be partially explained by methodological
397 biases because previous studies used traditional methodologies of stomach content analysis. Although
398 Romero et al. (2021) used genetic methods, they were only used to assist in the identification of hard
399 parts. When using stomach content analysis, stomachs that seem empty or have highly digested prey
400 are often discarded from the analysis, and depending on the experimental design, the majority of the
401 stomachs can be empty (e.g. 61% of empty stomachs in Ménard et al. 2000, 85% of empty stomachs
402 in Karakulak et al. 2009). The use of metabarcoding enables to overcome most of these problems and
403 biases, because the whole stomach content, including highly digested material, is used. Therefore,
404 not only it allows to extract prey DNA from empty stomachs that would have been discarded in
405 conventional studies, but also detects DNA of soft-bodied prey that would rapidly be digested (Diaz
406 Briz et al., 2017; Magnuson, 1969) and will not be detected using conventional stomach content
407 analysis. Inclusively, metabarcoding has recently revealed the importance of gelatinous prey on the

408 diet of Atlantic Bluefin tuna (*Thunnus thynnus*) (Günther et al., 2021), which would otherwise be
409 nearly impossible to detect.

410

411 4.2. Prey overlap with seabirds

412 Overall, when looking only to the values of the overlap index (all lower than 0.58), the prey
413 diversity overlap between seabird and tuna species was relatively low, especially between Skipjack
414 tuna and both seabird species. The greater frequency of krill (Euphausiidae) and siphonophores
415 (Diphyidae) on the diet of Skipjack tunas contributes greatly to this low overlap, while Brown boobies
416 and Cape Verde shearwaters prey mainly on epipelagic fish, and some of the fish species consumed
417 are of commercial interest (Almeida et al., 2021; Carreiro et al., 2022; Rodrigues, 2014; Vieira, 2018).
418 Another explanation for the low overlap between tunas and seabirds can be the higher prey
419 availability for tunas. Because some tuna species perform vertical migrations (Bertrand et al., 2002a),
420 they have a wider range of habitats to forage, which contributes to their higher prey diversity.
421 However, when we focus on prey from the epipelagic fraction of the ocean, that both tunas and
422 seabirds can explore, we observe a higher overlap of prey diversity.

423 On the other hand, from all prey diversity detected, some of the most abundant families were
424 shared between both tuna and seabird species. Of the ten families that Brown Boobies prey upon,
425 nine are shared with at least one of the tuna species. Indeed, this resulted in a significant overlap index
426 between Brown boobies and Yellowfin tuna, especially due to the higher frequency of Flying fish
427 (Exocoetidae) and Halfbeak fish (Hemiramphidae) in the diet of these predators. Carreiro et al.
428 (2022) analysed the diet of Brown boobies with a more comprehensive range of samples and showed
429 that these prey are the most important in their diet. Flying fishes inhabit tropical and subtropical
430 waters (Lewallen et al., 2017, 2016; Parin and Shakhovskoy, 2000) and historical fishery landings in
431 Cabo Verde indicate that flying fishes are present on Cabo Verde waters throughout the year
432 (Almeida, 2021). Both flying and halfbeak fishes are largely described as one of the main prey of
433 marine top predators in tropical regions, including tunas (da Silva et al., 2019; Vaske Júnior et al.,

434 2003) and seabirds (Ashmole and Ashmole, 1967; Correia et al., 2021; Mancini and Bugoni, 2014).
435 Furthermore, the fact that Yellowfins also prey on these two prey groups further hints that Brown
436 boobies in Cabo Verde may take advantage of facilitated foraging provided by tunas, at least from
437 Yellowfin tuna. Concerning temporal differences, as well as differences between Skipjack tuna
438 samples from August with the other groups, it is very likely that these were due to the patchy
439 distribution of prey and the foraging strategies of tunas, highly focused towards feeding on prey which
440 tends to aggregate in schools.

441

442 **5. Conclusions**

443 In Cabo Verde, we found that Skipjack tuna also explores prey of lower trophic levels, while
444 Yellowfin tuna prey mostly on mid-level trophic levels. It is well known that overfishing large
445 predatory fishes can have cascading effects through top-down perturbations (Casini et al., 2012;
446 Steneck, 2012; Verity et al., 2002). Besides local fisheries, countries like Spain, Japan, Taiwan, North
447 Korea and China operate in Cabo Verde waters, with or without fishing agreements (Kroodsma et al.,
448 2018). Because national authorities are not able to control all these foreign fleets, some countries may
449 use illegal and destructive fishing methods, which contribute to the decline of fish stocks and local
450 fish landings (Benchimol et al., 2009). Therefore, the results of this study can have implications for
451 the management of tuna fisheries in Cabo Verde. Regarding seabirds, we show that the overlap is
452 higher with tunas that prey on pelagic fish, namely the Yellowfin tuna. Although we cannot assert if
453 such overlap has negative effects on the viability of these seabirds populations, these results should
454 not be overlooked because the overexploitation of these tuna species may have a direct effect on the
455 availability of prey to seabirds.

456 Finally, the results of this study also confirm that ecosystem-based management is needed for
457 this study region, through the inclusion of other top predators such as seabirds. Because we detected
458 temporal differences, future research should focus on studying the diet of both tunas and seabirds
459 throughout the year, and include other seabirds that are present year-round in the archipelago and also

460 known to be a target of bycatch, such as the Red-billed tropicbird (*Phaethon aethereus*) (Montrond
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462

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482

483 **Declarations of interest:** none

484

485 Data Availability

486 Data analyzed during the current study are available publicly on Biostudies:
487 <https://www.ebi.ac.uk/biostudies/studies/S-BSST918>

488

489 Compliance with Ethical Standards

490 This project was authorized by the "National Directorate of the Environment" of Cabo Verde
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494

495 Author Contributions

496 Ana R. Carreiro: investigation; Ana Carreiro, Vitor H. Paiva, Jaime A. Ramos and Ricardo J.
497 Lopes: conceptualization, methodology, resources; Ana R. Carreiro and Ricardo J. Lopes: formal
498 analysis, data curation, writing - original draft; Vitor H. Paiva, Jaime A. Ramos and Ricardo J. Lopes:
499 funding acquisition, supervision, writing – review & editing; Vanessa A. Mata: software, writing –
500 review & editing; Nathalie M. Almeida, Isabel Rodrigues, Ivo dos Santos, Diana M. Matos, Pedro
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503

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- 815

816 **Tables**817 **Table 1** - Primers used for prey DNA screening and blocking primers designed for this study.

Genetic target	Taxa target	Pipeline/Aim	Primer set	Primer names	Sequence 5' - 3'	Amplicon size (bp)	Reference
mtDNA 12S	Osteichthyes	Screening	MiFish-U	MiFish-U-F	GTCGGTAAAACTCGTGCCAGC	165-185	Miya et al. 2015
				MiFish-U-R	CATAGTGGGGTATCTAATCCCAGTTTG		
	<i>Thunnus</i> sp.	Blocking	<i>Thunnus</i> 12S BP	MiFish-blkThunnus	CTAATCCCAGTTTGTGTCATAGCTTTCGTGGGGTCAGGGT[SpC3]	-	This study
mtDNA COI	<i>Katsuwonus</i> sp.	Blocking	<i>Katsuwonus</i> 12S BP	MiFish-blkKatsuwonus	CTAATCCCAGTTTGTGTCATAGCTTTCGTGGGGTCAGGGG[SpC3]	-	This study
	Metazoa	Screening	Leray-XT	mICOIintF-XT	GGWACWRGWTGRACWITITAYCCYCC	~313	Wangensteen et al. 2018
				jpgHCO2198	TAIACYTCIGGRTGICCRAARAAYCA		
	Scombridae	Blocking	Tuna COI BP	Leray-blkTuna	AAGAATCAGAATAGGTGTTGGTAAAG[SpC3]	-	This study

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821 **Table 2** – Results of the multivariate analysis of the prey MOTUs family occurrence matrix
 822 between species (CE = Cape Verde shearwater, SL = Brown booby, SKJ = Skipjack tuna, YFT =
 823 Yellowfin tuna) and between months (Jul = July, Aug = August, Sep = September). Only significant
 824 ($p < 0.05$) pairwise and univariate comparisons are presented.

Factor	p	Significant pairwise	p	Significant univariate	p		
		comparisons		comparisons			
Species	0.001	CE ≠ SL ≠ SKJ ≠ YFT	all < 0.003	Engraulidae	0.002		
				Exocoetidae	0.001		
				Diphyidae	0.002		
				Euphausiidae	0.001		
				Ommastrephidae	0.040		
				Scombridae	0.002		
Month	0.002	Aug ≠ Jul	0.002	Gonostomatidae	0.017		
				Aug ≠ Sep	0.031	Phosichthyidae	0.017
Species : Month	0.001	SKJ : Aug ≠ YFT : Jul	0.001	Carangidae	0.049		
				SKJ : Aug ≠ SL : Sep	0.012	Exocoetidae	0.037
				SKJ : Aug ≠ CE : Aug	0.027		
				SKJ : Aug ≠ SL : Jul	0.033		
				YFT : Jul ≠ YFT : Aug	0.027		
		YFT : Aug ≠ SL : Sep	0.022				

825

826 Figures

827 **Figure 1.** a) Location of the Cabo Verde Archipelago in the eastern Atlantic Ocean and its
 828 Exclusive Economic Zone (EEZ); b) The species highlighted in this article, as representative
 829 members of two groups of top marine predators (Seabirds and Tunas), likely to interact at the trophic
 830 level.

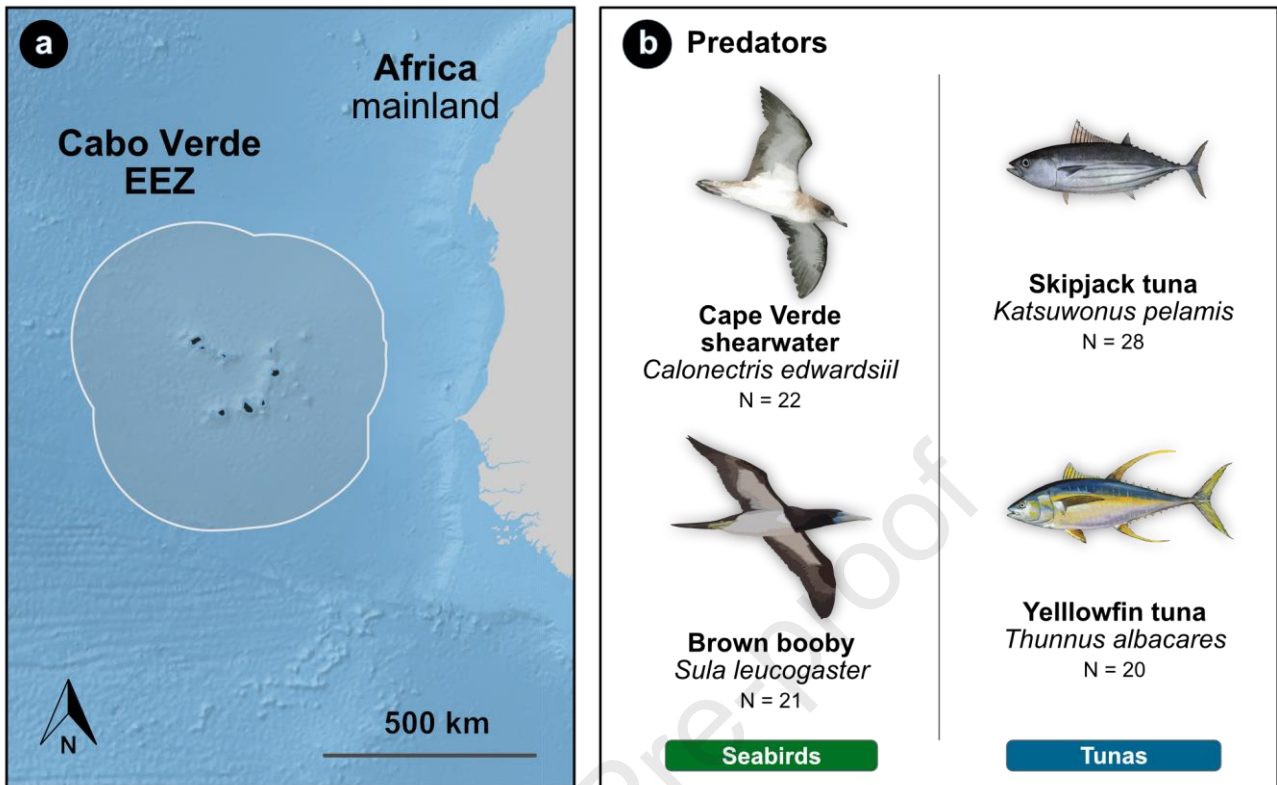
831 **Figure 2.** Visual characterization of the number of prey MOTUs families in the diet of the studied
 832 seabirds and tunas, ordered by taxonomic similarity. Colour classifies each MOTU family according
 833 to its frequency of occurrence.

834 **Figure 3.** a) Line plots of ranked frequency of occurrence, highlighting the most frequent
835 MOTUs families for each predator species; b) Pianka's niche overlap index for the tuna and seabird
836 predator species. Significant values ($p < 0.05$) are highlighted with an asterisk.

837 **Figure 4.** Non-metric multidimensional scaling ordination (nMDS) of prey MOTU family
838 composition (Jaccard dissimilarity) in the diet of the studied seabirds and tunas. Convex hull polygons
839 delineate the four predator species. The nMDS scores for all prey MOTUs families are shown as
840 points. The name of the prey MOTUs family is only shown when their frequency of occurrence is
841 high ($>20\%$).

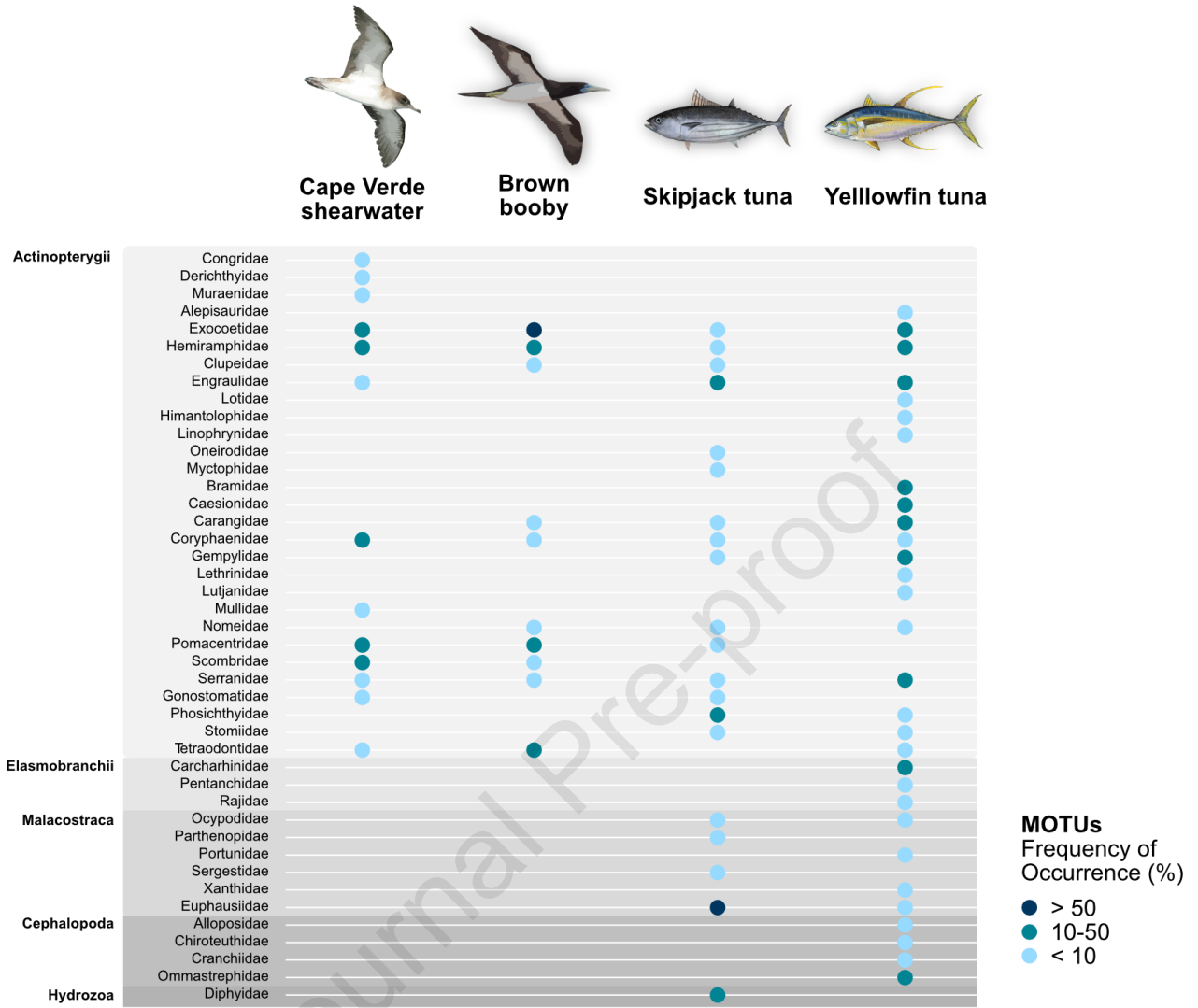
842

Journal Pre-proof

843 **Figure 1**

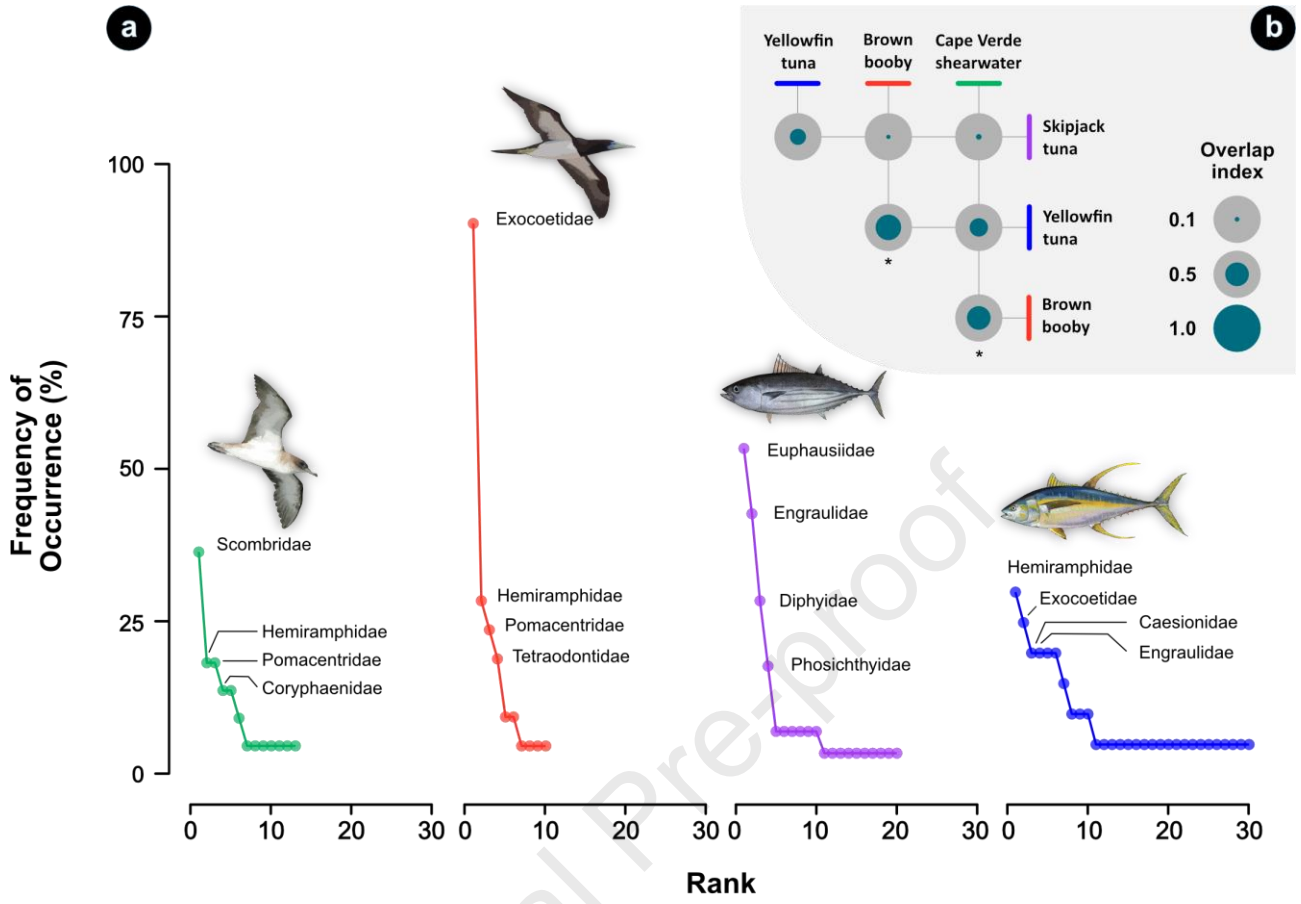
844

845 **Figure 2**



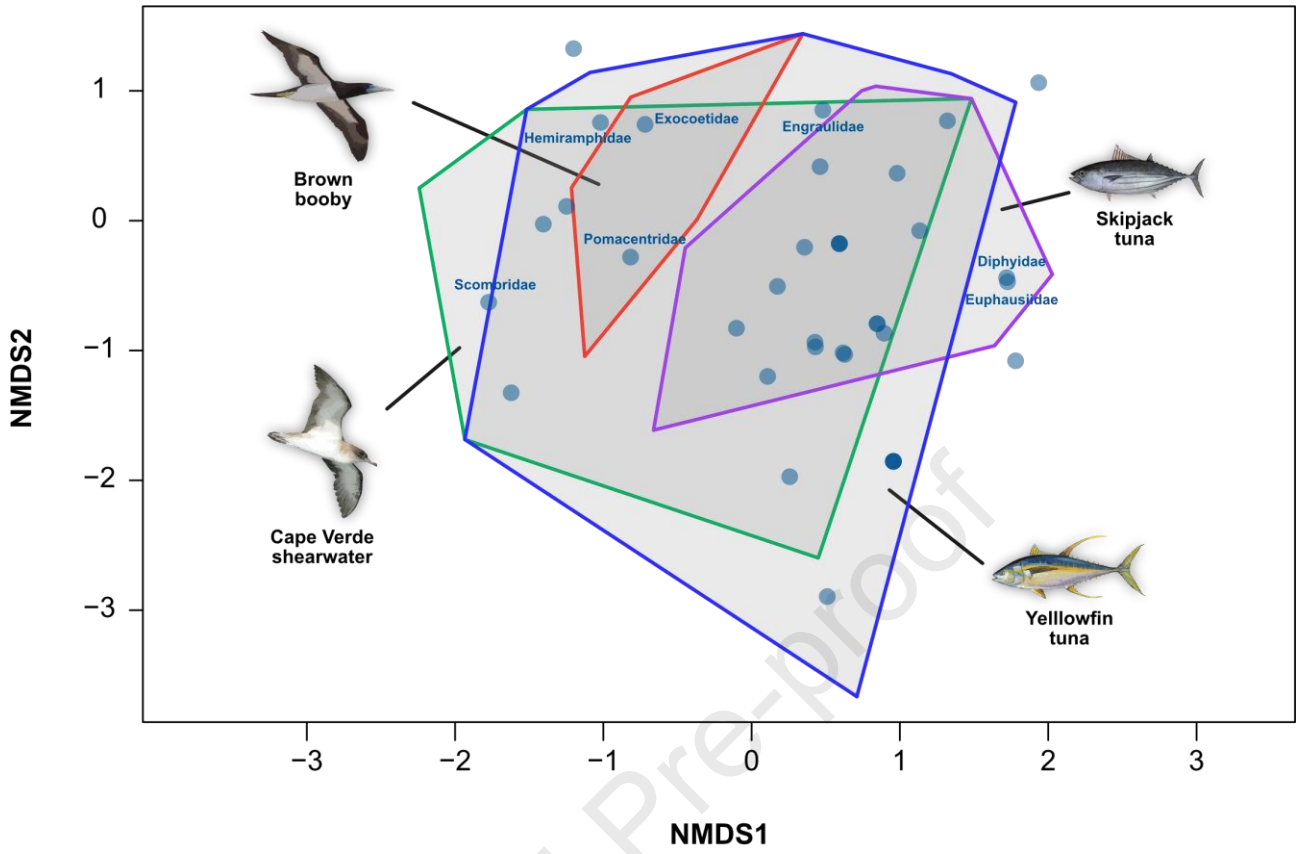
846

847 **Figure 3**



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849 **Figure 4**



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852 **Appendices**853 **Appendix A.** PCR conditions for each of the HTS libraries.

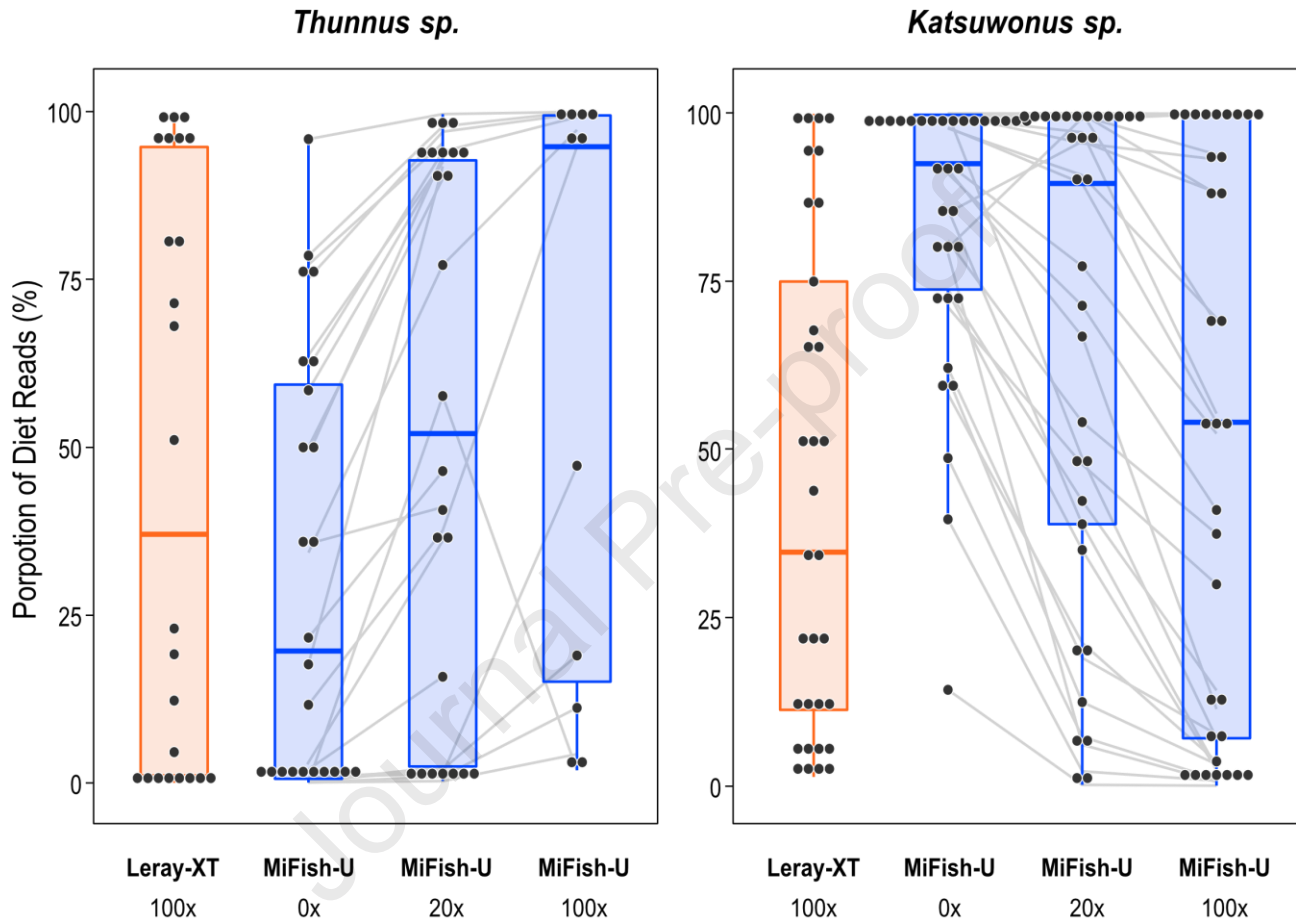
Library	Blocking primer concentration	Volume of reaction (μl)	Number of cycles	Denaturing	Annealing	Extension
Tuna	MiFish-U 0X	10	35	94°C - 30s	60° - 30s	72° - 60s
	MiFish-U 20X	12	35	94°C - 30s	60° - 30s	72° - 60s
	MiFish-U 100X	12	35	94°C - 30s	60° - 30s	72° - 60s
	Leray-XT 100X	12	35	94°C - 30s	45° - 30s	72° - 60s
Seabird	MiFish-U 0X	10	35	94° - 30s	60° - 30s	72° - 60s
	Leray-XT 0X	10	40	94° - 30s	45° - 45s	72° - 60s

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856 **Appendix B.** Boxplots representing the proportion of diet reads per sample. Differences
857 between libraries are shown for each of the blocking primers designed, for *Thunnus* sp. samples (left
858 panel) and *Katsuwonus* sp. samples (right panel). Different concentrations of blocking primer were
859 connected with a grey line to visualize the trend for each sample between MiFish-U libraries.

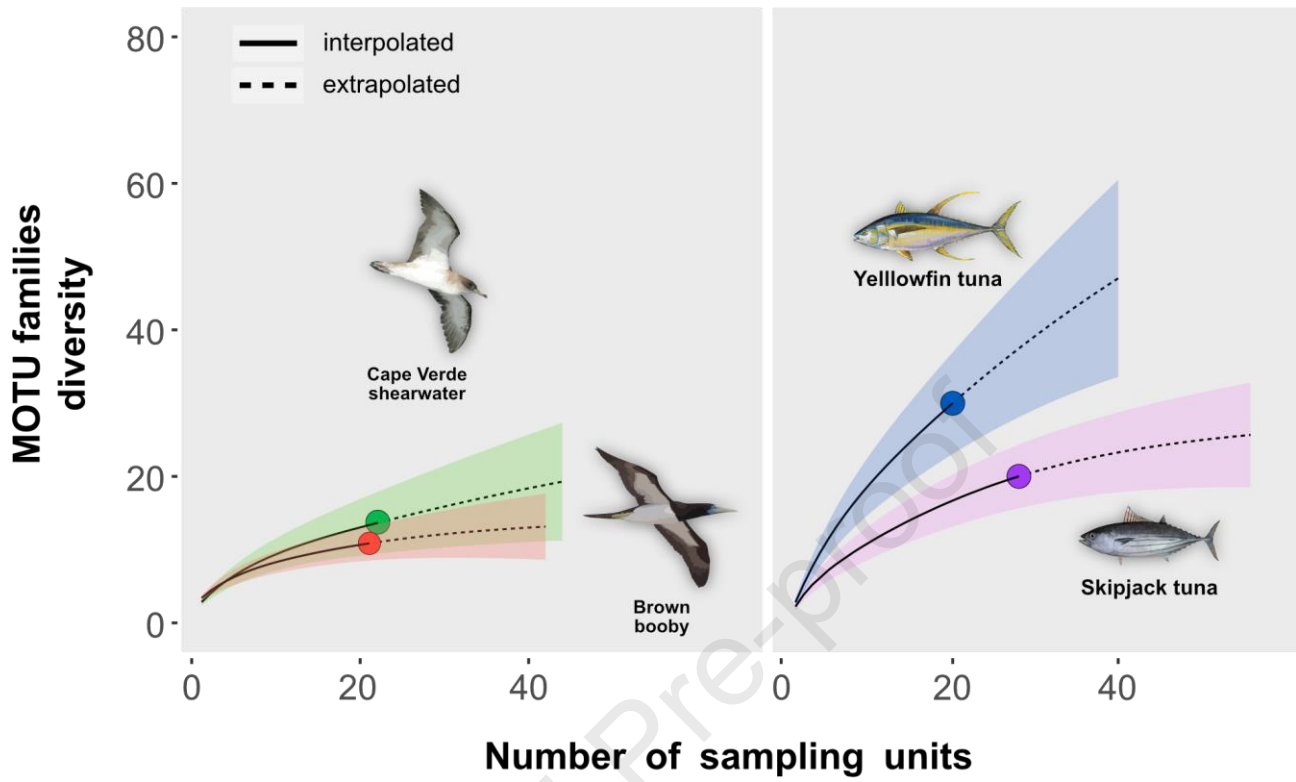
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863 **Appendix C.** Prey MOTUs family's accumulation curves and extrapolated richness, with 95%
864 confidence intervals, for each predator species.



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866

1 **DNA metabarcoding to assess prey overlap between tuna and seabirds in the**
2 **Eastern Tropical Atlantic: implications for an ecosystem-based management**

3

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20 **Highlights**

- 21 • DNA metabarcoding assess diet of tuna and seabirds of Eastern Atlantic Ocean.
22 • Skipjack tuna diet was dominated by prey from lower trophic levels.
23 • Significant diet overlap between Brown booby and Yellowfin tuna.
24 • Results have implications for the management of fisheries in this region.

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2 **Eastern Tropical Atlantic: implications for an ecosystem-based management**

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20 **Author Contributions**

21 Ana R. Carreiro: investigation; Ana Carreiro, Vitor H. Paiva, Jaime A. Ramos and Ricardo
22 J. Lopes: conceptualization, methodology, resources; Ana R. Carreiro and Ricardo J. Lopes:
23 formal analysis, data curation, writing - original draft; Vitor H. Paiva, Jaime A. Ramos and
24 Ricardo J. Lopes: funding acquisition, supervision, writing – review & editing; Vanessa A. Mata:
25 software, writing – review & editing; Nathalie M. Almeida, Isabel Rodrigues, Ivo dos Santos,
26 Diana M. Matos, Pedro M. Araújo, Teresa Militão, Jacob González-Sólis: resources, writing –
27 review & editing. All authors read and approved the final manuscript.

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20 **Declaration of interest:** None

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